

# The influence of binaural effects on annoyance for transportation noise

Jaehwan Kim,<sup>a)</sup> Changwoo Lim,<sup>b)</sup> Jiyoung Hong,<sup>c)</sup> Wontae Jung<sup>d)</sup> and Soogab Lee<sup>e)</sup>

(Received 2006 March 17; revised 2006 December 27; accepted 2007 January 15)

**A laboratory study was conducted to identify the significance of the binaural effect on annoyance by transportation noise. There were 377 participants (240 male and 137 female). Every participant marked the annoyance score (from 0 to 10) for each stimulus on their questionnaire. All of data were divided into four groups according to four types of transportation noise sources. The difference in the mean annoyance scores between subgroups according to sound recording/playback techniques was statistically significant in the four noise groups. Stepwise multiple regression analysis was conducted. Using a regression model, predictions of  $L_{Aeq}$ , Technique and their interaction term ( $L_{Aeq}$ \*Technique) involving dummy variable were made. In spite of some limitations of the process of the laboratory test, meaningful results were acquired. This study showed that the binaural effect was one of the acoustical factors modifying annoyance, and quantitative expression of binaural effect is suggested. Annoyance was explained by  $L_{Aeq}$  and an interaction term in the resultant model functions for the four noise sources. Regression coefficients of the interaction term for each model equation were almost the same, which explained the extent of the binaural effect. The binaural effect was defined as the difference of annoyance variations between two sub-groups when  $L_{Aeq}$  varies by  $\Delta L_{Aeq}$  from the specified noise level. The binaural effect was expressed as  $0.01\Delta L_{Aeq}$  on the basis of a 0 to 10 annoyance scale. © 2007 Institute of Noise Control Engineering.**

Primary subject classification: 63.2; Secondary subject classification: 52.3

## 1 INTRODUCTION

Environmental noises, such as transportation noise, recreational noise, industrial noise and community noise are recognized as environmental pollutants. Therefore, many studies on noise assessment have been performed in many countries in the field<sup>1-5</sup> and in the laboratory.<sup>6-12</sup> Thanks to these dedicated efforts, it is known that not only acoustical but non-acoustical factors influence annoyance. In particular, laboratory

studies continue to report acoustical and non-acoustical factors that influence annoyance by comparing annoyance responses in the different situations.

Rylander et al.<sup>6</sup> selected noise from heavy trucks as stimuli in a laboratory experiment. Heavy truck noise included background noise from road traffic caused by other vehicles in a normal traffic situation. Also, window attenuation was simulated by a filter that reduced noise by 5 dB per octave. Subjects were tested in two different conditions (different noise level with the same number of noise events and different number of noise events with the same noise level). The results were that the extent of annoyance increased with the increase in  $L_{Aeq}$  when the number of noise events was equal. At the same  $L_{Aeq}$ , the increase in the number of noise events caused an increase in annoyance.<sup>6</sup> Rasmussen<sup>7</sup> reported a relation between the number of events and subjective value caused by three noise sources and mixed stimuli. Relations between annoyance and noise metrics such as continuous equivalent sound pressure level ( $L_{eq}$ ), noise pollution level ( $LNP$ ) and three measures based on time-derivative, were compared.<sup>7</sup> According to Rasmussen, three measures based on time-derivative may be interpreted as  $L_{eq}$  plus

<sup>a)</sup> Department of Mechanical and Aerospace Engineering, Seoul National University, Seoul 151-744, KOREA; e-mail: kjh03@snu.ac.kr.

<sup>b)</sup> BK21 School for Creative Engineering Design of Next Generation Mechanical and Aerospace Systems, Seoul National University, Seoul 151-744, KOREA.

<sup>c)</sup> Department of Mechanical and Aerospace Engineering, Seoul National University, Seoul 151-744, KOREA.

<sup>d)</sup> Department of Mechanical and Aerospace Engineering, Seoul National University, Seoul 151-744, KOREA. Presently at Applied Technology, Hyundai MOBIS.

<sup>e)</sup> Center for Environmental Noise and Vibration Research, Department of Mechanical and Aerospace Engineering, Seoul National University, Seoul 151-744, KOREA; e-mail: solee@plaza.snu.ac.kr.

Table 1—Acoustical characteristics and noise level of 15 noise events in test set 1.

No. of test set	No.	Noise source	Measurement device	$L_{Aeq}$ (dB)
1	1	Air(commercial)	Conventional microphone	65
	2	Road traffic	Dummy head	80
	3	Air(military)	Conventional microphone	40
	4	Air(commercial)	Dummy head	50
	5	Road traffic	Conventional microphone	55
	6	Air(military)	Dummy head	80
	7	Railway vehicle	Conventional microphone	85
	8	Road traffic	Dummy head	60
	9	Air(commercial)	Dummy head	80
	10	Road traffic	Conventional microphone	94
	11	Air(military)	Conventional microphone	75
	12	Air(military)	Dummy head	60
	13	Railway vehicle	Conventional microphone	70
	14	Railway vehicle	Dummy head	40
	15	Road traffic	Dummy head	65

a correction term which should correspond to the expected increase in annoyance when the sound pressure level fluctuates. Also, research on annoyance by difference noise sources with various frequency characteristics has been performed. The difference of annoyance response caused by different noise sources was shown.<sup>8,10</sup>

In most research on community noise annoyance and noise assessment, noise metrics as independent variables for explanation of annoyance response are calculated from single microphone signals. Human being's auditory organ is, however, physiologically a

two channel input system and sound reception is affected by the pinna, head and body as sound enters each ear channel.<sup>13</sup> The external ears distort the original sounds and the extent of distortion is represented by the external ear transfer functions (usually called Head Related Transfer Functions (HRTFs), which depends on the distance and direction from source related to the head).<sup>13</sup> The external ear is composed of pinna, head and torso. The acoustical effects of the pinna are based on reflection, diffraction, shadowing and dispersion.<sup>13</sup> Dominant acoustical effects around the head vary according to wave length (compared with head dimen-

Table 2—Acoustical characteristics and noise level of 14 noise events in test set 2.

No. of test set	No.	Noise source	Measurement device	$L_{Aeq}$ (dB)
2	1	Air(military)	Conventional microphone	70
	2	Road traffic	Conventional microphone	40
	3	Air(commercial)	Dummy head	75
	4	Air(military)	Dummy head	55
	5	Railway vehicle	Conventional microphone	45
	6	Railway vehicle	Dummy head	60
	7	Air(military)	Conventional microphone	85
	8	Air(commercial)	Dummy head	90
	9	Air(military)	Conventional microphone	90
	10	Road traffic	Dummy head	85
	11	Railway vehicle	Conventional microphone	75
	12	Air(military)	Conventional microphone	45
	13	Air(commercial)	Dummy head	55
	14	Railway vehicle	Dummy head	65

Table 3—Acoustical characteristics and noise level of 15 noise events in test set 3.

No. of test set	No.	Noise source	Measurement device	$L_{Aeq}$ (dB)
3	1	Road traffic	Conventional microphone	70
	2	Air(commercial)	Dummy head	70
	3	Railway vehicle	Conventional microphone	55
	4	Air(military)	Dummy head	50
	5	Air(commercial)	Conventional microphone	45
	6	Road traffic	Dummy head	90
	7	Air(commercial)	Conventional microphone	85
	8	Air(commercial)	Dummy head	60
	9	Railway vehicle	Dummy head	80
	10	Railway vehicle	Conventional microphone	88
	11	Road traffic	Conventional microphone	75
	12	Road traffic	Dummy head	50
	13	Railway vehicle	Conventional microphone	50
	14	Air(commercial)	Dummy head	40
	15	Air(military)	Dummy head	65

sion). Because of these acoustical effects around head and pinna, the ears play an important role of spatial hearing.<sup>13</sup> Because of the HRTFs, noise received by a dummy head has different frequency and level characteristics relative to the same signal received by a single microphone. This means that sounds used in calculating noise metrics are different from sounds heard by subjects. But, no comment on this fact is included in the above-mentioned papers. Through the headphone reproduction of the single microphone signals and HATS (Head And Torso Simulator, Brüel & Kjær Type 4100) signal, subjective responses to the signals recorded by a single microphone and the binaural signal recorded by a dummy head are compared in this

paper. There were some limitations to this laboratory study. Each listener's head is unique. Each individual HRTF is slightly different from HRTF of HATS. When subjects listen to test sounds, headphone corrections (including a correction for sound field between headphone and subjects' ears) for individual ears are slightly different from that in the artificial ears of HATS. But, most dummy heads have been designed using an averaging technique so that their HRTFs are the average of individual HRTFs.<sup>13</sup> These effects are less important than the fact that tonal auditory signals are distorted by the HRTFs.<sup>13</sup> In this paper, it will be proposed and demonstrated by laboratory study that the

Table 4—Acoustical characteristics and noise level of 14 noise events in test set 4.

No. of test set	No.	Noise source	Measurement device	$L_{Aeq}$ (dB)
4	1	Railway vehicle	Conventional microphone	65
	2	Air(commercial)	Conventional microphone	55
	3	Air(military)	Dummy head	45
	4	Railway vehicle	Dummy head	75
	5	Road traffic	Conventional microphone	85
	6	Air(military)	Dummy head	90
	7	Air(commercial)	Conventional microphone	90
	8	Air(military)	Dummy head	85
	9	Railway vehicle	Conventional microphone	60
	10	Railway vehicle	Dummy head	45
	11	Air(military)	Conventional microphone	55
	12	Air(commercial)	Conventional microphone	75
	13	Road traffic	Dummy head	40
	14	Air(military)	Dummy head	70

Table 5—Acoustical characteristics and noise level of 15 noise events in test set 5.

No. of test set	No.	Noise source	Measurement device	$L_{Aeq}$ (dB)
5	1	Air(military)	Conventional microphone	65
	2	Air(commercial)	Conventional microphone	40
	3	Railway vehicle	Dummy head	50
	4	Road traffic	Conventional microphone	50
	5	Road traffic	Dummy head	75
	6	Railway vehicle	Dummy head	88
	7	Railway vehicle	Conventional microphone	80
	8	Air(commercial)	Conventional microphone	60
	9	Air(commercial)	Dummy head	85
	10	Road traffic	Conventional microphone	90
	11	Air(commercial)	Dummy head	45
	12	Air(military)	Conventional microphone	50
	13	Railway vehicle	Dummy head	55
	14	Air(commercial)	Conventional microphone	70
	15	Road traffic	Dummy head	70

binaural effect is one of the acoustic factors that have an effect on annoyance.

## 2 MEASUREMENT AND STIMULUS

### 2.1 Measurement

In order to investigate the binaural effects on noise annoyance, transportation noise was recorded by single microphone (Brüel & Kjør 4190) and by HATS (Head And Torso Simulator, Brüel & Kjør Type 4100) simultaneously. Both measurement devices were placed about 15 m away from the edge of road and railway, and about 100 m away from takeoff and landing point

of military and commercial aircraft at a 1.7 m height above the ground. The single microphone was located about 5 m away from the dummy head along a line parallel to the track of noise sources. It is impossible for two measurement devices to be located at the same position simultaneously and it is unavoidable that dummy head obstructs the sound. However, noises reaching the two measurement devices are almost same because measurements were performed in open environment with no obstacles at nearly equal distance from the source. Also, the extent of acoustical interference of the signal measured by the microphone

Table 6—Acoustical characteristics and noise level of 15 noise events in test set 6.

No. of test set	No.	Noise source	Measurement device	$L_{Aeq}$ (dB)
6	1	Road traffic	Conventional microphone	65
	2	Railway vehicle	Conventional microphone	40
	3	Railway vehicle	Dummy head	70
	4	Air(military)	Conventional microphone	60
	5	Air(military)	Dummy head	75
	6	Road traffic	Dummy head	94
	7	Air(commercial)	Conventional microphone	80
	8	Road traffic	Conventional microphone	60
	9	Railway vehicle	Dummy head	85
	10	Air(military)	Conventional microphone	80
	11	Road traffic	Dummy head	55
	12	Air(commercial)	Conventional microphone	50
	13	Air(military)	Dummy head	40
	14	Road traffic	Conventional microphone	80
	15	Air(commercial)	Dummy head	65

Table 7—Number of participants according sex in each test set.

No. of test set	Set 1	Set 2	Set 3	Set 4	Set 5	Set 6
Male	39	37	32	36	47	49
Female	20	22	26	26	21	22
Total	59	59	58	62	68	71

produced by the dummy head was negligible, because the dummy head was placed about 5 m away from the microphone.

The two measurement devices were directly connected to a PULSE (Brüel & Kjær 3560C) for sound recording, real-time monitoring and spectral analysis. Each recorded sound from commercial aircraft, military aircraft and railway was a single noise event, which was generated from one passby. Recorded road traffic noise represented by general highway traffic contained a number of vehicles (heavy and light vehicles) whose velocity ranged from 80 to 120 km/h. Sampling rate was set to 65536 Hz with 16 bit quantization.

## 2.2 Constitution of Stimuli

Using Cool edit pro ver. 2.0 software, military and commercial aircraft noises were generated with the

increment of  $L_{Aeq}$  (equivalent continuous A-weighted sound pressure level) of 5 dBA from  $L_{Aeq}=40$  to 90 dBA. Railway and road traffic noise events were generated for 40, 45, 50, 55, 60, 65, 70, 75, 80, 85 and 88 dBA and 40, 50, 55, 60, 65, 70, 75, 80, 85, 90 and 94 dBA, respectively. A total of 6 test sets were composed from 88 noise events. 44 noise events were generated from the sounds that had been recorded by the single microphone and the others were generated from the sounds that had been recorded by the dummy head. 88 noise events were randomly arranged in each test set. The constitution of each test set is summarized in Tables 1–6. The duration of the recordings of the aircraft noise, railway noise and road traffic noise events were about 15 sec. Silence between noise events in six test sets lasted about 10 sec.

## 3 LABORATORY EXPERIMENT

### 3.1 Subjects

Total 380 persons participated in laboratory study. One person's data was lost and two of the 380 persons did not have normal hearing (i.e. hearing loss greater the 20 dB of RETSPL<sup>14</sup>). All of subjects were paid fees for participation. Available data was acquired from tests conducted with 377 subjects (240 men and 137

Table 8—Relation between equivalent noise level and annoyance from commercial aircraft noise according to participants' sex.

$L_{Aeq}$ (dBA)	Measurement device	Annoyance (mean $\pm$ standard deviation)		
		male	female	average
40	Conventional microphone	2.36 $\pm$ 1.67	1.62 $\pm$ 1.77	2.13 $\pm$ 1.73
45	Conventional microphone	2.81 $\pm$ 1.75	3.27 $\pm$ 2.31	3.02 $\pm$ 2.01
50	Conventional microphone	2.78 $\pm$ 1.79	2.73 $\pm$ 1.55	2.76 $\pm$ 1.71
55	Conventional microphone	4.14 $\pm$ 1.97	3.69 $\pm$ 4.14	3.95 $\pm$ 1.97
60	Conventional microphone	4.06 $\pm$ 1.72	4.14 $\pm$ 2.03	4.09 $\pm$ 1.81
65	Conventional microphone	4.15 $\pm$ 2.40	3.50 $\pm$ 1.99	3.93 $\pm$ 2.27
70	Conventional microphone	5.94 $\pm$ 1.89	6.19 $\pm$ 1.17	6.01 $\pm$ 1.70
75	Conventional microphone	6.47 $\pm$ 2.14	6.81 $\pm$ 1.96	6.61 $\pm$ 2.06
80	Conventional microphone	7.20 $\pm$ 2.15	7.27 $\pm$ 1.58	7.23 $\pm$ 1.98
85	Conventional microphone	8.56 $\pm$ 1.68	8.42 $\pm$ 1.68	8.50 $\pm$ 1.67
90	Conventional microphone	8.97 $\pm$ 1.63	9.08 $\pm$ 1.57	9.02 $\pm$ 1.59
40	Dummy head	2.91 $\pm$ 2.13	3.00 $\pm$ 2.04	2.91 $\pm$ 2.13
45	Dummy head	3.04 $\pm$ 1.61	2.81 $\pm$ 1.89	3.04 $\pm$ 1.61
50	Dummy head	4.18 $\pm$ 2.43	3.15 $\pm$ 2.16	4.18 $\pm$ 2.43
55	Dummy head	4.30 $\pm$ 2.20	4.05 $\pm$ 2.03	4.30 $\pm$ 2.20
60	Dummy head	4.91 $\pm$ 1.94	4.81 $\pm$ 2.08	4.91 $\pm$ 1.94
65	Dummy head	5.69 $\pm$ 2.27	6.18 $\pm$ 1.82	5.69 $\pm$ 2.27
70	Dummy head	6.75 $\pm$ 1.98	6.38 $\pm$ 2.26	6.75 $\pm$ 1.98
75	Dummy head	7.92 $\pm$ 2.31	7.91 $\pm$ 2.18	7.92 $\pm$ 2.31
80	Dummy head	8.33 $\pm$ 2.34	8.30 $\pm$ 1.87	8.33 $\pm$ 2.34
85	Dummy head	9.04 $\pm$ 1.23	9.43 $\pm$ 0.87	9.04 $\pm$ 1.23
90	Dummy head	9.76 $\pm$ 0.64	9.32 $\pm$ 1.62	9.76 $\pm$ 0.64

Table 9—Relation between equivalent noise level and annoyance from military aircraft noise according to participants' sex.

$L_{Aeq}$ (dBA)	Measurement device	Annoyance (mean $\pm$ standard deviation)		
		male	female	average
40	Conventional microphone	1.79 $\pm$ 1.59	2.35 $\pm$ 1.95	1.98 $\pm$ 1.73
45	Conventional microphone	2.27 $\pm$ 1.63	2.14 $\pm$ 1.78	2.22 $\pm$ 1.67
50	Conventional microphone	3.02 $\pm$ 1.66	2.48 $\pm$ 1.86	2.85 $\pm$ 1.73
55	Conventional microphone	3.33 $\pm$ 1.66	3.58 $\pm$ 2.47	3.44 $\pm$ 2.02
60	Conventional microphone	4.49 $\pm$ 1.67	3.59 $\pm$ 2.22	4.21 $\pm$ 1.89
65	Conventional microphone	4.72 $\pm$ 1.78	3.81 $\pm$ 1.57	4.44 $\pm$ 1.76
70	Conventional microphone	5.84 $\pm$ 1.80	4.77 $\pm$ 2.02	5.44 $\pm$ 1.94
75	Conventional microphone	5.54 $\pm$ 2.19	5.45 $\pm$ 1.39	5.51 $\pm$ 1.94
80	Conventional microphone	6.69 $\pm$ 2.15	6.95 $\pm$ 1.99	6.77 $\pm$ 2.09
85	Conventional microphone	8.92 $\pm$ 1.42	8.14 $\pm$ 1.91	8.63 $\pm$ 1.65
90	Conventional microphone	8.76 $\pm$ 1.64	8.00 $\pm$ 2.00	8.47 $\pm$ 1.80
40	Dummy head	2.06 $\pm$ 1.59	2.09 $\pm$ 1.87	2.07 $\pm$ 1.67
45	Dummy head	3.25 $\pm$ 2.32	2.58 $\pm$ 1.79	2.97 $\pm$ 2.13
50	Dummy head	3.50 $\pm$ 1.81	3.77 $\pm$ 2.12	3.62 $\pm$ 1.95
55	Dummy head	4.19 $\pm$ 1.93	3.82 $\pm$ 2.22	4.05 $\pm$ 2.03
60	Dummy head	4.03 $\pm$ 1.99	3.90 $\pm$ 1.62	3.98 $\pm$ 1.86
65	Dummy head	5.97 $\pm$ 2.16	6.15 $\pm$ 2.41	6.05 $\pm$ 2.26
70	Dummy head	6.25 $\pm$ 2.14	6.73 $\pm$ 2.16	6.45 $\pm$ 2.15
75	Dummy head	7.22 $\pm$ 2.09	6.91 $\pm$ 1.66	7.13 $\pm$ 1.96
80	Dummy head	7.79 $\pm$ 2.26	8.55 $\pm$ 1.99	8.05 $\pm$ 2.18
85	Dummy head	8.33 $\pm$ 1.76	8.54 $\pm$ 1.79	8.42 $\pm$ 1.76
90	Dummy head	9.11 $\pm$ 1.62	9.46 $\pm$ 1.17	9.26 $\pm$ 1.45

women). Their ages ranged from 20 to 61 years. All of subjects heard only one test set among six test sets in an anechoic chamber after taking an audiometric screening test. They marked the extent of annoyance from each noise event on a questionnaire.

### 3.2 Apparatus

In order to test subjective response to transportation noise, a system which could control the sound in an anechoic chamber through a desk top PC (Pentium IV) with DIGI 96 Pro sound card (RME) was constructed. Only the subject and headphone (Sennheiser HD25) were located in the anechoic chamber. In addition, sound monitoring system was connected to a PULSE (Brüel & Kjær Type 3560C) to calibrate the sound pressure level and spectrum level before each laboratory test.

Frequency characteristics of the single microphone as well as microphones mounted in the ears of the dummy head, which were used to record transportation noise events, were flat in the frequency range from 20 Hz to 20 kHz. So, it could be said that there was no distortion of the recorded sounds. However, frequency response of headphone and sound field between headphone and subjects' ears could distort the original

sounds. In order to compensate for the sound distortion caused by these two problems, white noise was heard by both ears of the dummy head through a headphone. Output terminal of the dummy head was directly connected to the PULSE, which contained a frequency analyzer and 1/3 octave band analyzer. The input signal to the dummy head ears (output sound for calibration) was analyzed in frequency domain. The inversion of the frequency response functions for the headphone and diffuse field (sound field between headphone and subjects' ears) provided correction factors which compensated for levels to which the subjects were exposed from recordings from the single microphone and the dummy head.

After the calibration of level and spectrum had been completed, the laboratory experiment was conducted in the anechoic chamber in order to avoid interference from other noises. The size of test section was 3.2  $\times$  3.2  $\times$  2.1 m and cut off frequency was approximately 200 Hz. Background noise levels in an anechoic chamber were about 20 ~ 23 dBA.

### 3.3 Procedures

All of the people who participated in laboratory test were audiometrically screened before the main test. If

Table 10—Relation between equivalent noise level and annoyance from railway noise according to participants' sex.

$L_{Aeq}$ (dBA)	Measurement device	Annoyance (mean $\pm$ standard deviation)		
		male	female	average
40	Conventional microphone	2.02 $\pm$ 1.55	1.55 $\pm$ 1.18	2.02 $\pm$ 1.55
45	Conventional microphone	2.73 $\pm$ 1.79	2.45 $\pm$ 1.95	2.73 $\pm$ 1.79
50	Conventional microphone	2.56 $\pm$ 1.61	3.00 $\pm$ 1.81	2.56 $\pm$ 1.61
55	Conventional microphone	2.91 $\pm$ 1.67	3.58 $\pm$ 2.21	2.91 $\pm$ 1.67
60	Conventional microphone	3.50 $\pm$ 2.01	3.69 $\pm$ 2.54	3.50 $\pm$ 2.01
65	Conventional microphone	4.25 $\pm$ 2.45	4.04 $\pm$ 1.91	4.25 $\pm$ 2.45
70	Conventional microphone	5.49 $\pm$ 2.19	5.20 $\pm$ 1.99	5.49 $\pm$ 2.19
75	Conventional microphone	6.16 $\pm$ 1.91	5.27 $\pm$ 2.29	6.16 $\pm$ 1.91
80	Conventional microphone	7.77 $\pm$ 1.54	7.43 $\pm$ 1.83	7.77 $\pm$ 1.54
85	Conventional microphone	8.74 $\pm$ 1.67	8.10 $\pm$ 2.63	8.74 $\pm$ 1.67
88	Conventional microphone	8.56 $\pm$ 1.63	8.46 $\pm$ 2.08	8.56 $\pm$ 1.63
40	Dummy head	1.82 $\pm$ 1.78	1.50 $\pm$ 1.43	1.71 $\pm$ 1.66
45	Dummy head	2.19 $\pm$ 1.47	2.54 $\pm$ 1.86	2.34 $\pm$ 1.64
50	Dummy head	3.72 $\pm$ 1.80	2.86 $\pm$ 2.03	3.46 $\pm$ 1.90
55	Dummy head	4.04 $\pm$ 1.99	3.71 $\pm$ 1.82	3.94 $\pm$ 1.93
60	Dummy head	5.03 $\pm$ 1.98	4.77 $\pm$ 2.33	4.93 $\pm$ 2.10
65	Dummy head	5.62 $\pm$ 2.11	5.00 $\pm$ 1.80	5.39 $\pm$ 2.01
70	Dummy head	6.02 $\pm$ 1.80	5.68 $\pm$ 1.78	5.92 $\pm$ 1.79
75	Dummy head	7.39 $\pm$ 2.11	7.69 $\pm$ 2.00	7.52 $\pm$ 2.05
80	Dummy head	7.91 $\pm$ 1.67	8.08 $\pm$ 2.19	7.98 $\pm$ 1.91
85	Dummy head	8.57 $\pm$ 1.95	8.73 $\pm$ 1.28	8.62 $\pm$ 1.76
88	Dummy head	9.17 $\pm$ 1.58	9.43 $\pm$ 1.21	9.25 $\pm$ 1.47

subjects had hearing losses less than 20 dB of RETSPL,<sup>14</sup> they were regarded as normal. After audiometric screening test, subjects were instructed on the procedure and the method of experiment. The main contents of the instructions were as follows.

- (1) Whatever makes a noise or disturbs an experiment should be ignored from subjects
- (2) Annoyance is different from level and loudness of noise events. Thus, level and loudness itself of each noise event must not be compared.
- (3) Subjects may imagine reading a book, watching TV or a similar activity.
- (4) Marking an annoyance score on the questionnaire should be performed when no sound is heard after hearing each noise event. Enough time to mark the annoyance score is given.

After instructions were given to the subjects, laboratory experiments followed. All of the subjects performed one test set among the 6 test sets. Subjects were exposed to both transportation noise recorded by the single microphone and by the dummy head in one test set. They then recorded the extent of annoyance in the questionnaire. Subjects were asked to answer the question, "What extent of annoyance do you feel as if you heard the noise in your common environment?"

There exists various annoyance scales such as 5- or 7- or 9-point verbal and numerical scale. However, annoyance rating scale adopted in this experiment was an 11-point numerical scale, ranging from zero to ten. Ten means "extremely annoyed" and zero means "not annoyed at all." The choice of the 11-point numerical scale is based on the assumption that respondents are more cognitively familiar with 0-10 scaling than with the shorter 7 or 9-point numeric scales.<sup>15</sup>

### 3.4 Status of Laboratory Test

The number of subjects for each test set is summarized in Table 7. 59 subjects performed test set 1 and 2. 58, 62, 68 and 71 subjects performed test set 3, 4, 5 and 6, respectively. Thus, 5534 points were collected in the database.

## 4 ANALYSIS AND RESULT

Tables 8–11 show the relation between annoyance and equivalent noise level from commercial aircraft, military aircraft noise, railway noise and road traffic noise. It was found from Tables 8–11 that the relation between noise levels and their corresponding annoyance score was almost linear. Also, the mean annoyance score from noise recorded by the dummy head

Table 11—Relation between equivalent noise level and annoyance from road noise according to participants' sex.

$L_{Aeq}$ (dBA)	Measurement device	Annoyance (mean $\pm$ standard deviation)		
		male	female	average
40	Conventional microphone	1.27 $\pm$ 1.04	1.68 $\pm$ 1.81	1.42 $\pm$ 1.38
50	Conventional microphone	2.47 $\pm$ 1.52	1.71 $\pm$ 1.45	2.24 $\pm$ 1.53
55	Conventional microphone	2.77 $\pm$ 2.07	2.75 $\pm$ 2.24	2.76 $\pm$ 2.11
60	Conventional microphone	2.76 $\pm$ 1.61	2.50 $\pm$ 1.47	2.68 $\pm$ 1.57
65	Conventional microphone	3.53 $\pm$ 1.92	2.82 $\pm$ 2.08	3.31 $\pm$ 1.98
70	Conventional microphone	4.50 $\pm$ 1.65	4.38 $\pm$ 1.75	4.45 $\pm$ 1.68
75	Conventional microphone	5.03 $\pm$ 1.73	4.96 $\pm$ 2.55	5.00 $\pm$ 2.12
80	Conventional microphone	6.94 $\pm$ 1.80	6.95 $\pm$ 1.84	6.94 $\pm$ 1.80
85	Conventional microphone	7.75 $\pm$ 2.14	8.19 $\pm$ 1.83	7.94 $\pm$ 2.02
90	Conventional microphone	7.98 $\pm$ 1.79	7.52 $\pm$ 2.04	7.84 $\pm$ 1.87
94	Conventional microphone	8.95 $\pm$ 1.41	8.20 $\pm$ 2.44	8.69 $\pm$ 1.84
40	Dummy head	1.78 $\pm$ 1.42	1.69 $\pm$ 1.89	1.74 $\pm$ 1.62
50	Dummy head	2.28 $\pm$ 1.75	2.19 $\pm$ 1.72	2.24 $\pm$ 1.72
55	Dummy head	2.24 $\pm$ 1.51	1.86 $\pm$ 1.25	2.13 $\pm$ 1.43
60	Dummy head	3.41 $\pm$ 2.10	2.85 $\pm$ 1.79	3.22 $\pm$ 2.00
65	Dummy head	4.92 $\pm$ 2.31	4.15 $\pm$ 1.95	4.66 $\pm$ 2.21
70	Dummy head	6.06 $\pm$ 1.76	6.14 $\pm$ 1.31	6.09 $\pm$ 1.63
75	Dummy head	7.26 $\pm$ 1.69	7.10 $\pm$ 1.64	7.21 $\pm$ 1.66
80	Dummy head	7.18 $\pm$ 2.32	7.55 $\pm$ 2.42	7.31 $\pm$ 2.34
85	Dummy head	8.51 $\pm$ 1.54	7.09 $\pm$ 2.43	7.98 $\pm$ 2.02
90	Dummy head	9.00 $\pm$ 2.02	9.15 $\pm$ 1.43	9.07 $\pm$ 1.77
94	Dummy head	9.22 $\pm$ 1.65	9.50 $\pm$ 1.01	9.31 $\pm$ 1.48

Table 12—Summary of the result of one-way ANOVA in commercial aircraft noise group.

	Sum of squares	Degrees of freedom	Mean square	F	Significance
Between groups	252.871	1	25.871	28.409	<0.0001
Within groups	122292.205	1381	8.901		
Total	12545.076	1382			

$F(1, \infty, 0.95)=3.84$ ;  $F(1, \infty, 0.99)=6.63$

Table 13—Robust tests of equality of means in military aircraft noise group.

	Statistic	Degrees of freedom 1	Degrees of freedom 2	Significance
Welch	21.731	1	1365.393	<0.0001
Brown-Forsythe	21.731	1	1365.393	<0.0001



Table 14—Summery of the results of one-way ANOVA in railway noise group.

	Sum of squares	Degrees of freedom	Mean square	F	Significance
Between groups	171.375	1	171.375	18.640	<0.0001
Within groups	12650.710	1376	9.194		
Total	12822.084	1377			

$$F(1, \infty, 0.95)=3.84; F(1, \infty, 0.99)=6.63$$

was slightly higher than that by the single microphone. In order to identify the statistical significance of binaural effects on annoyance more precisely, all data were divided into four groups according to the four noise sources. Statistical analysis (one-way ANOVA and regression analysis) was conducted for the four groups.

As is well known, ANOVA is a method to compare means between two or more groups. In this research, means of two categories (or sub-groups) classified by sound recording/playback technique were compared in four groups. Homogeneous variance assumption should be met for which the result of ANOVA was reliable. Preliminary test, which is Levene's test, for homogeneity of variance, was carried out in each noise source group. Variances in the two categories (according to sound recording/playback technique) were equal ( $P > 0.05$ ) in the cases of commercial aircraft noise and railway noise but were not in the other cases ( $P < 0.05$ ). If the equal variances assumption has been violated, the Welch or Brown-Forsythe statistic instead of the usual  $F$ -statistic was used to test for the difference of means between groups.<sup>16</sup> Tables 12 and 14 show the results of ANOVA for the commercial aircraft

noise and the railway noise group. Tables 13 and 15 show the results of robust tests for equality of means for the military aircraft noise group and the road traffic noise group. As shown in Tables 12–15, the significance of  $F$  statistic and Welch statistic (or Brown-Forsythe statistic) was less than 0.01. Therefore null hypothesis (meaning that the two groups are equal) was rejected in all cases. It was statistically significant that binaural effects influenced the annoyance ( $p < 0.01$ ).

Stepwise multiple regression analysis was conducted in each group in order to examine more precisely that binaural effect was an effective modifier. The annoyance score was the response variable. Let  $A$ -weighted equivalent level, dummy variable (sound recording/playback technique) and their interaction term ( $L_{Aeq} * Technique$ ) involving the dummy variable be predictors in order that the interaction effect as well as main effect could be considered. Sound recording/playback technique as predictor will be simply referred to as the technique from henceforth. Properties of the dummy variable and interaction term ( $L_{Aeq} * Technique$ ) are shown below.

$$\begin{aligned} Technique &= 0 \text{ in case of the microphone recording} \\ &= 1 \text{ in case of the dummy head recording} \end{aligned} \quad (1)$$

$$\begin{aligned} Interaction Term &= 0 \text{ in case of the microphone recording} \\ &= LAeq \text{ in case of the dummy head recording} \end{aligned} \quad (2)$$

The basic procedure of stepwise regression analysis is as follows. Firstly, the independent variable best

correlated with the dependent variable is included in the regression equation. The significance of the

Table 15—Robust tests of equality of means in road traffic noise group.

	Statistic	Degrees of freedom 1	Degrees of freedom 2	Significance
Welch	18.229	1	1384.106	<0.001
Brown-Forsythe	18.229	1	1384.106	<0.001

included independent variable is checked. If it is not significant, regression analysis is stopped and regression model is determined to be the mean value of response variable. Otherwise, the independent variable with the highest partial *F* with the dependent variable among the remaining independent variables is entered. This process was repeated until the addition of a remaining independent variable did not increase R-squared by a significant amount (significance level used in this research is 0.05).<sup>17</sup>

The result of stepwise regression analysis is summarized in Table 16. In this table, group A, B, C and D means the commercial aircraft noise, military aircraft

noise, railway noise and road traffic noise, respectively. It was found that regression coefficients were statistically significant ( $p < 0.01$ ) and the entry of the interaction term ( $L_{Aeq} * Technique$ ) caused a significant increase in the R-square value in Table 16. In all groups, predictors of the resultant model equations were A-weighted equivalent level and interaction term ( $L_{Aeq} * Technique$ ) so that the mathematical expression of resultant regression model was as below.

$$\hat{Y}(\text{annoyance}) = b_0 + b_1 L_{Aeq} + b_2 L_{Aeq} * Technique \quad (3)$$

$$\begin{aligned} \hat{Y}(\text{annoyance}) &= b_0 + b_1 L_{Aeq} \quad \text{for } Technique = 0 \\ &= b_0 + (b_1 + b_2) L_{Aeq} \quad \text{for } Technique = 1 \end{aligned} \quad (4)$$

Table 16—Summary of the results of regression analysis in each group (regression models and coefficients<sup>c</sup>).

Group	Model	Unstandardized coefficients		Standardized coefficients	t	Significance	R	R Square	Adjusted R Square
		Std. error	Beta						
A	1 (Intercept)	-3.734	.226		-16.515	<0.0001			
	$L_{Aeq}$	.144	.003	.752	42.427	<0.0001	.752 <sup>a</sup>	.566	.566
	2 (Intercept)	-3.721	.221		-16.814	<0.0001			
B	$L_{Aeq}$	.137	.003	.720	40.330	<0.0001	.765 <sup>b</sup>	.585	.584
	$L_{Aeq} \times Technique$	.012	.002	.140	7.869	<0.0001			
	1 (Intercept)	-3.878	.224		-17.327	<0.0001			
C	$L_{Aeq}$	.141	.003	.750	41.989	<0.0001	.750 <sup>a</sup>	.562	.562
	2 (Intercept)	-3.869	.220		-17.610	<0.0001			
	$L_{Aeq}$	.135	.003	.719	39.897	<0.0001	.760 <sup>b</sup>	.578	.577
D	$L_{Aeq} \times Technique$	.011	.002	.131	7.269	<0.0001			
	1 (Intercept)	-4.540	.223		-20.362	<0.0001			
	$L_{Aeq}$	.151	.003	.773	45.187	<0.0001	.773 <sup>a</sup>	.597	.597
E	2 (Intercept)	-4.519	.220		-20.537	<0.0001			
	$L_{Aeq}$	.146	.003	.746	42.815	<0.0001	.780 <sup>b</sup>	.608	.608
	$L_{Aeq} \times Technique$	.010	.002	.108	6.200	<0.0001			
F	1 (Intercept)	-5.601	.226		-24.764	<0.0001			
	$L_{Aeq}$	.155	.003	.796	49.027	<0.0001	.796 <sup>a</sup>	.633	.633
	2 (Intercept)	-5.592	.222		-25.186	<0.0001			
G	$L_{Aeq}$	.150	.003	.768	46.980	<0.0001	.804 <sup>b</sup>	.647	.646
	$L_{Aeq} \times Technique$	.010	.001	.120	7.333	<0.0001			

<sup>a</sup>Predictors: (Intercept),  $L_{Aeq}$

<sup>b</sup>Predictors: (Intercept),  $L_{Aeq}$ ,  $L_{Aeq} \times Technique$

<sup>c</sup>Dependent variable: Annoyance

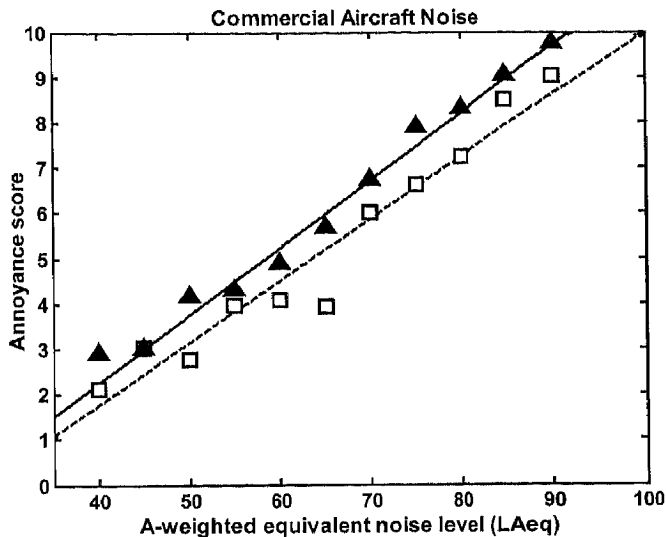


Fig. 1—Comparison of subjective responses to commercial aircraft noise between two different sound recording/playback techniques (Real line and triangular symbol represents regression model and observed data when subjects were tested through sound recorded by a dummy head, respectively. Dotted line and rectangular symbol represents regression model and observed data when subjects were tested through sound recorded by a single microphone, respectively).

This means that the regression models for the two categories (sound recording by the single microphone and by the dummy head) have the same intercept but different slopes for all four noise sources. Figures 1–5 show the differences in the annoyance responses according to sound recording/playback technique for each noise group. The extent of annoyance to the sounds recorded with the dummy head was higher than that to the sounds recorded with the single microphone. It could be seen from Eq. (4) that the difference in the annoyance scores between two categories is given by the term  $b_2 L_{Aeq}$ . It was expected that this term would reflect binaural effects. Moreover,  $b_2$ , which was the regression coefficient of interaction term, was similar for each noise source;  $b_2$  is 0.012 in commercial aircraft noise group (A), 0.011 in military aircraft noise group (B) and 0.01 in railway and road traffic noise groups (C, D).

Let binaural effects be defined as the difference in the variations in the annoyance scores between two categories when  $L_{Aeq}$  is varied by  $\Delta L_{Aeq}$  from the specified noise level. For example, when  $L_{Aeq}$  was increased by  $\Delta L_{Aeq}$  from the specified noise level, the binaural effect caused an increase in the annoyance response by

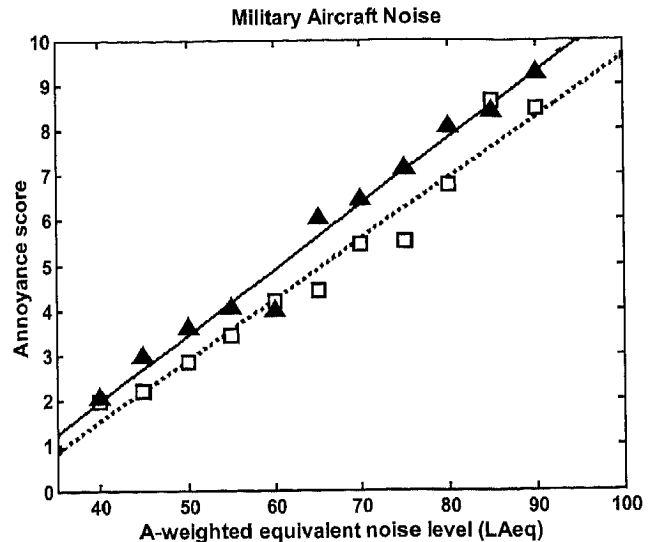


Fig. 2—Comparison of subjective responses to military aircraft noise between two different sound recording/playback techniques (Real line and triangular symbol represents regression model and observed data when subjects were tested through sound recorded by a dummy head, respectively. Dotted line and rectangular symbol represents regression model and observed data when subjects were tested through sound recorded by a single microphone, respectively).

as much as an additional  $0.01 \Delta L_{Aeq}$ . Figure 5 explains in more detail the binaural effects (difference between  $\Delta A1$  and  $\Delta A0$ ) defined in this paper. Note that the binaural effects explained in this paper are based on from 0 to 10 annoyance scale.

## 5 CONCLUDING REMARKS

Noise measurements have been carried out with a single microphone and noise metrics have been calculated based on input signals from the single microphone. However, people hear sound that has been distorted by binaural effects. This fact means that subjective response graphs (noise levels vs. annoyance responses) represent the relation between noise levels and their corresponding annoyance responses plus binaural effects.

In this paper, the binaural effect was defined as not a physical quantity but as a psychological (or subjective) quantity. For the purpose of demonstrating the significance of binaural effect and evaluating the binaural effect as a psychological quantity, a laboratory study was conducted using headphone simulation in the

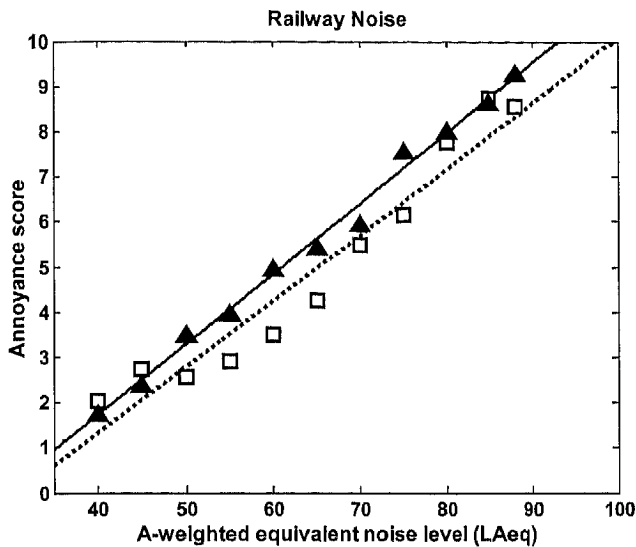


Fig. 3—Comparison of subjective responses to railway noise between two different sound recording/playback techniques (Real line and triangular symbol represents regression model and observed data when subjects were tested through sound recorded by a dummy head, respectively. Dotted line and rectangular symbol represents regression model and observed data when subjects were tested through sound recorded by a single microphone, respectively).

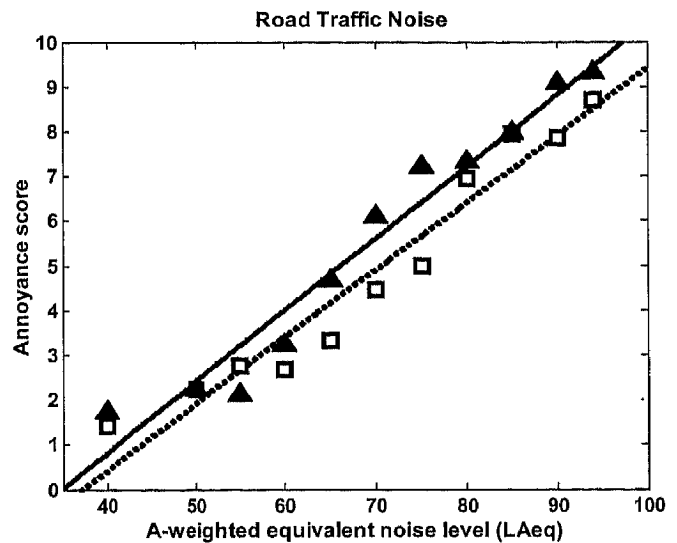


Fig. 4—Comparison of subjective responses to road traffic noise between two different sound recording/playback techniques (Real line and triangular symbol represents regression model and observed data when subjects were tested through sound recorded by a dummy head, respectively. Dotted line and rectangular symbol represents regression model and observed data when subjects were tested through sound recorded by a single microphone, respectively).

anechoic chamber. Of course, there were some limitations in these tests. Except for prior mentioned things (see Secs. 1 and 2.1), sound attenuation with propagation in each frequency band was not considered in detail.

However, important results could be acquired from laboratory tests in spite of these limitations. For the four noise sources, it was found that the mean annoyance scores for signals recorded with a single microphone and with a dummy head were significantly different. Binaural effects were defined as the difference of annoyance variations between two sub-groups when  $L_{Aeq}$  was varied by  $\Delta L_{Aeq}$  from the specified noise level. The predictors were  $L_{Aeq}$  and  $L_{Aeq} * Technique$  in resultant regression of four noise sources, where *Technique* is zero for recording with the single microphone and one for recording with the dummy head. The regression coefficients of  $L_{Aeq} * Technique$  in four model equations for the four transportation noise sources explained the extent of binaural effect. In the four model equations showed a value for  $b_2=0.01$ , indicating that annoyance increased faster by about 1% with than without the binaural effect. Binaural effect

could be expressed as  $0.01\Delta L_{Aeq}$  on the basis of from 0 to 10 annoyance scale.

However, it would be too early to conclude that binaural effect was  $0.01\Delta L_{Aeq}$ . It is necessary that the quantity of the binaural effect be verified through more studies in more realistic environments.

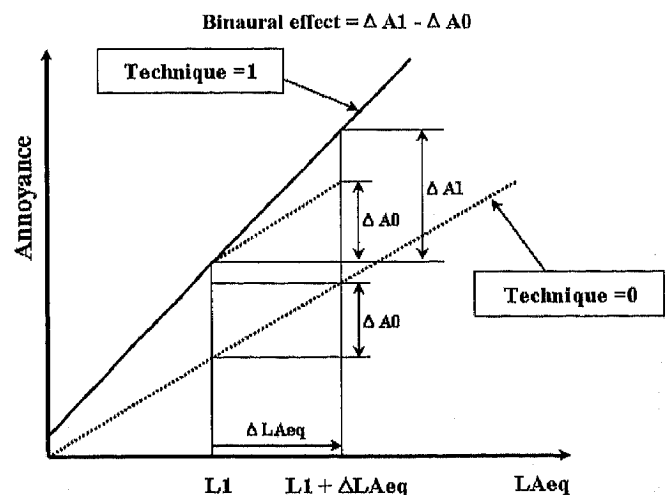


Fig. 5—Geometric description of binaural effects defined in this paper.

## 6 ACKNOWLEDGMENTS

This work was supported by the 'Core Environmental Technology Development Project for Next Generation in Korea Institute of Environmental Science and Technology', and 'Brain Korea 21 Project in 2006'.

## 7 REFERENCES

1. T. J. Schultz, "Synthesis of social surveys on noise annoyance," *J. Acoust. Soc. Am.* **64**(2), 377-405 (1978).
2. S. Fidell, D. S. Barber and T. J. Schultz, "Updating a dosage-effect relationship for the prevalence of annoyance due to general transportation noise," *J. Acoust. Soc. Am.* **89**(1), 221-233 (1991).
3. L. S. Finegold, C. S. Harris and H. E. von Gierke, "Community annoyance and sleep disturbance: Updated criteria for assessing the impacts of general transportation noise on people," *Noise Control Eng. J.* **42**(1), 25-30 (1994).
4. Henk M. E. Miedema and Henk Vos, "Exposure-response relationships for transportation noise," *J. Acoust. Soc. Am.* **104**(6), 3432-3445 (1998).
5. Henk M. E. Miedema and Henk Vos, "Demographic and attitudinal factors that modify annoyance from transportation noise," *J. Acoust. Soc. Am.* **105**(6), 3336-3344 (1999).
6. R. Rylander, E. Sjöstedt, and M. Björkman, "Laboratory studies on traffic noise annoyance," *J. Sound Vib.* **52**(3), 415-421 (1977).
7. K. B. Rasmussen, "Annoyance from simulated road traffic noise," *J. Sound Vib.* **65**(2), 203-214 (1979).
8. E. Öhström, M. Björkman and R. Rylander, "Laboratory annoyance and different traffic noise sources," *J. Sound Vib.* **70**(3), 334-341 (1980).
9. G. Labiale, "Laboratory study of the influence of noise level and vehicle number on annoyance," *J. Sound Vib.* **90**(3), 361-371 (1983).
10. N. J. Versfeld and J. Vos, "Annoyance caused by sounds of wheeled and tracked vehicles," *J. Acoust. Soc. Am.* **101**(5), 2677-2685 (1977).
11. N. J. Versfeld and J. Vos, "A-weighted equivalent sound level as a predictor of the annoyance caused by road traffic consisting of various proportions of light and heavy vehicles," *J. Sound Vib.* **253**(2), 389-399 (2002).
12. G. M. Aasvang and B. Engdahl, "Subjective responses to aircraft noise in an outdoor recreational setting: a combined field and laboratory study," *J. Sound Vib.* **276**(3-5), 981-996 (2004).
13. J. Blauert, *Spatial Hearing: The Psychophysics of Human Sound Localization*, MIT Press, Cambridge, MA (1983).
14. International Organization for Standardization ISO 389-1: 1998(E), Acoustics - Reference zero for the calibration of audiometric equipment - Part 1: Reference equivalent threshold sound pressure level for pure tones and supra-aural earphone (1998).
15. International Standard Organization ISO/TS 15666, Acoustics - assessment of noise annoyance by means of social and socio-acoustic surveys (2003).
16. H. Lee and Y. Kim, Spss 10.0 Manual in Korean, Bobmunsa (2003).
17. Norman R. Draper and Harry Smith, *Applied regression analysis*, 3rd edition, Wiley Series (1998).